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Microorganism Communities Response of Ecological Changes in Post Tin Mining Ponds

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Abstract

The ecological changes are the main problem of post tin mining activities, such as post tin mining ponds and it has occurred for a long time chronosequence. The understanding of how post tin mining activities bring ecological changes and how the microorganism communities respond to these environment changes becomes an important part to explore the potential of bioremediator to accelerate the recovery of water quality as water sources for secondary activities and habitable for organisms and further, to determine the next aquatic ecosystem management. The change of ecological factors such as pH, dissolved oxygen (DO), and heavy metals can be focused for a discussion as affecting factors for changes in microorganism communities. In fact, all of the three factors can drive the alteration of microecosystem and indirectly causes change in microorganism communities. However, the survival capability of microorganisms, resistance and resilience capacity, and the interaction of microorganisms with ecosystem change, in particular, in post tin mining ponds chronosequence has not been explained properly.

Keywords: post tin mining, ecological change, chronosequence, microorganisms

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INTRODUCTION

Tin mining was one of mining activities that contribute to heavy metals contamination and environmental problems. In Heipang District of BarkinLadi of Nigeria, heavy metals were detected with very high concentrations of Pb, Zn, Mn, Fe, Cr, and Cu and also Ni in surface soil, soil, and sediment [1]; in the sediments of former tin mining catchment Bestari Jaya of Peninsular Malaysia was detected Sn, Pb, Zn, Cr, Cu, and As [2]; in post tin mining of Bangka Island of Indonesia was detected Fe, Al, Pb, Cu, and Zn [3, 4].

Furthermore, tin mining activities can also cause ecological problems such as temperature change, land degradation consequence, structure and composition of sand, clay, nitrogen, organic carbon, pH, cation exchange capacity (CEC), metal, etc [5–8]. Indirectly tin mining has changed the ecosystem and organism communities structure including an impact on microclimate in microscopic habitats and microorganisms's life [9].

Tin mining activity in land area was done by soil excavation; and soil formation was the

result of a complex network of biological process as well as chemical and physical processes. The ecological consequence of tin mining activity was a degradation of soil composition, structure, quality, and physical or biological characteristics, erosion damages to land use or land cover, and changes of macro and microorganisms in their natural habitats [10–14]. The changes of microorganism communities in their habitat could have significant effect on ecosystem function [15] and also on abundance, diversity, and productivity of macro communities living there [16–18].

Post tin mining microecosystem with nutrients deficiencies, acidic pH, low dissolved oxygen (DO), metal oxidizing and the other extreme environmental factors can be uninhabitable for most microorganisms. However, few studies investigated the microorganism's life during ecological change process [19] or microbial succession pattern in change through time [20]. Few studies on post tin mining, in particular, only investigated diversity of methane-oxidizing bacteria (MOB) and ammonia-oxidizing bacteria (AOB) in disused

ponds of post tin mining [21, 22]. While, an understanding of microorganisms and their activity as biological indicator of the contaminated ecosystem was crucial in order to predict ecosystem conditions and detect environment changes [23, 24] because they have a capability to respond quickly to the alterations, including in soil and aquatic ecosystem [25, 17].

Therefore, this review aims to gather information about the ecological changes that occur as a result of mining activities, especially in disused pond of post tin mining and to explain how the microorganism communities respond to the environmental changes caused by the tin mining chronosequence. The understanding of microbial adaptation capacity in post tin mining, particularly, in disused ponds of post tin mining and microorganisms' succession pattern for a long time effect of chronosequence becomes an important part to determine the next aquatic ecosystem management.

ECOLOGICAL CHANGE AND MICROORGANISM RESPONSE

Ecological Changes of Post Tin Mining

The increasing tin mining activity has produced various wastes such as tailings, oil, and fuel coming from the sand scraper tin boat [26]. Furthermore, heavy metals were detected from tin mining locations. In Malaysia, the tin mining activity has led to heavy metals contamination such as arsenic, copper, lead, tin and zinc in soil and then they were transferred to plants and it also happened for two shallow water aquatic plants [27] and heavy metals of tin mining was accumulated in different tissues and organs of organism, such as fish [28]. In the sediments of former tin mining catchment Bestari Jaya, Peninsular Malaysia indicated that the sediments have been polluted with arsenic (8.8%), chromium (12.9%), copper (17.4%), lead (19.5%), zinc (14.9%) and tin (33.8%) [2].

In additional to heavy metal, the characteristics of top soil samples from post tin mining area, Bestari Jaya, Peninsular Malaysia showed pH values ranging from acidic to neutral (4.8–7.2) [29]. The soils of post tin mining land generally had low CEC,

organic matter, total nitrogen, available phosphorus, macronutrient and clay content in soil texture. The recovery of this condition to natural characteristics can be done by natural succession; however, development of natural succession even takes a very long time [30]. The effects of post tin mining contaminations also take place in aquatic ecosystem such as marine, river, and disused or reused post tin mining ponds.

The study of water quality and heavy metals in water of post mining area showed that there was variation in the environment [31]. Each water body has a pattern of physical and chemical characteristics which were determined largely by the climatic, geomorphological, and geochemical conditions prevailing in the drainage basin and the underlying aquifer [32]. As a reference, in Bangka Island of Indonesia, the study results showed contaminations of Fe was 2.48–8.17 mg/l, Al was 0.04–3.34 mg/l, Cu was < 0.002–0.14 mg/l, and Zn < 0.007–0.450 mg/l in Kolong Hijau (green pond). In Kolong TB 1.9 (TB pond 1.9), contaminations of Fe, Al, Cu, and Zn were 2,230–32,270 mg/l, 7,560–80,000 mg/l, 0,0–0,080 mg/l, dan 0,0–0,870 mg/l [4]. North Central Nigeria, also experienced similar problems of heavy metals contamination from tin mining. A manganese value of 0.9 mg/l which was higher than the World Health Organization (WHO) highest desirable level of 0.05 mg/l was recorded from a mine pond and chromium values of 0.1 mg/l and 0.12 mg/l, respectively, which exceeded the maximum admissible concentration of 0.005 mg/l [33].

Microorganism Response of Post Tin Mining Ecosystem

The changes in post tin mining ecosystems characteristic as well as in other ecosystems certainly followed the changes of the organisms including the structure of microbial communities [34]. Microorganisms like bacteria can adapt to various environmental factors by modifying their membranes, in particular, phospholipid fatty acids composition and interaction between proteins and lipids [35]. Phospholipids are essential membrane components that make up a relatively constant proportion of the microorganisms under natural conditions. Their

patterns provide an insight into the structure of microorganism communities and biomass so they are used as one of the useful tool identifying microorganisms and characterizing microbial communities in natural systems [36].

Some factors, particularly, in aquatic ecosystem of post tin mining, can be focused for attention which might significantly influence the life of microorganisms there. As stated earlier, it is important to know the response of microorganisms to the ecological changes in post tin mining aquatic ecosystem and the response of microorganisms shown by their diversity and metabolic activity [37]. In principle, the succession of microorganisms in post tin mining have a similarity with the succession of microorganisms, generally, in others post mining.

However, the critical factors such as pH value, DO, and type of heavy metals can be focused as indicators in chronosequence study of microbial communities' response and succession.

Effect of pH Value on Microorganism Communities

pH value has been shown to be an important factor driving microbial communities [38]. As a reference, pH value of post tin mining water in Bangka Island in TB 1.9 (young post tin mining pond) was 2.21–3.42 and in Kolong Hijau (old post tin mining pond) was 4.82–6.48 [4]. pH value of post tin mining water was less than 7; therefore, post tin mining water can be characterized as acidic water.

Acidic pH and low DO are the characteristics of metal and sand mining activity [31]. Acidic waste waters from industrial and mining activities as acid mine drainage (AMD) waters were often highly acidic (pH < 4) [39]. Acid mine drainage was a result from the oxidization of sulfide minerals (principally pyrite, FeS₂), primarily via microbial-mediated reactions, after the mine wastes were exposed to oxygen and water [40]. Acidic, metal-rich waters generated by the microbial accelerated dissolution of pyrite and other sulfide minerals were frequently encountered in derelict mine

sites, including many that have been long-abandoned [41].

Acidic water can eliminate alkaline-tolerant microorganism and some microorganism in neutral pH. However, acid-tolerant microorganism can survive and even microorganisms that have a pH optimum for growth of acidic pH termed as acidophiles consists of extreme acidophiles (pH 2.7 or pH < 3) and moderate acidophiles (pH 5.7 or pH 3–5) [42, 43].

Acidophiles can accelerate dissolution of pyrite and other sulfide minerals in tailings as mineral waste generated by metal mining act as sources of acid [44]. They can generate energy by the conversion process of ferrous to ferric iron in acidic environment [45], and are psychrotolerant obligate autotrophs that appears to use only ferrous iron as an electron donor and oxygen as an electron acceptor [46].

To survive in acidic pH, acidophiles maintain a pH gradient of several pH units across the cellular membrane while producing ATP by the influx of protons through the F₀F₁ ATPase [47]. Some researchers have shown that pH value causes changes in microorganism (e.g. bacteria) composition, abundance, or diversity that Firmicutes, Actinobacteria, Gamma-Proteobacteria tend to dominant at pH 2.7, Acidobacteria are very dominant at pH 3 and pH 5.5, Firmicutes also dominant at pH 5.7, while Gamma-Proteobacteria are very dominant in alkaline condition (pH 10.3) (Figure 1).

Other studies also represented the effect of pH on diversity of microorganism communities [48]. The changes in structure of microorganism communities may occur in post tin mining pond with acidic to neutral pH value; however, a study is required to prove the same.

Effect of DO Value on Microorganism Communities

In addition to acid pH value, low DO selects microorganisms that can survive in the environment. In surface tailings water, communities detected were an unusual mixture of obligate aerobes, obligate anaerobes, and

facultative anaerobes [51]. In post tin mining activity, MOB (the γ -proteobacterial and α -proteobacteria) and AOB were isolated from disused tin mining lakes [21, 22].

The microorganisms in post tin mining ponds can be identified as facultative anaerobic microorganisms because they can use oxygen for their activity such as methane and ammonia oxidation, but also may grow in the environment without oxygen or low oxygen concentration. In fact, the oxygen-depleted water of mining was dominated by acidophilic communities with chemolithotrophic facultative anaerobe properties [52]; the chemolithotrophic facultative anaerobe microorganism usually dominates all natural and man-made acidic environments. Acidophilic species has a mixotrophy strategy,

as heterotrophic that uses organic compounds in the form of dissolved organic carbon (DOC) (chemoheterotroph) and as autotrophic that uses light, inorganic carbon, and mineral nutrients for photosynthesis (photoautotroph) [45].

The DO and DOC concentration can drive the microorganism (e.g. bacteria) diversity change. As a reference, the results in Figure 2 show that *Aciditibacillus*, *Leptospirillum*, and *Ferroplasma* dominate in DO 3, 8 and 5; while *Ferroplasma* and *Acidiphilium* very dominant in DOC 6, 8 or low DO. The community structure change may also occur in post tin mining ponds with low oxygen concentration; although, a study needs to be undertaken for a justification.

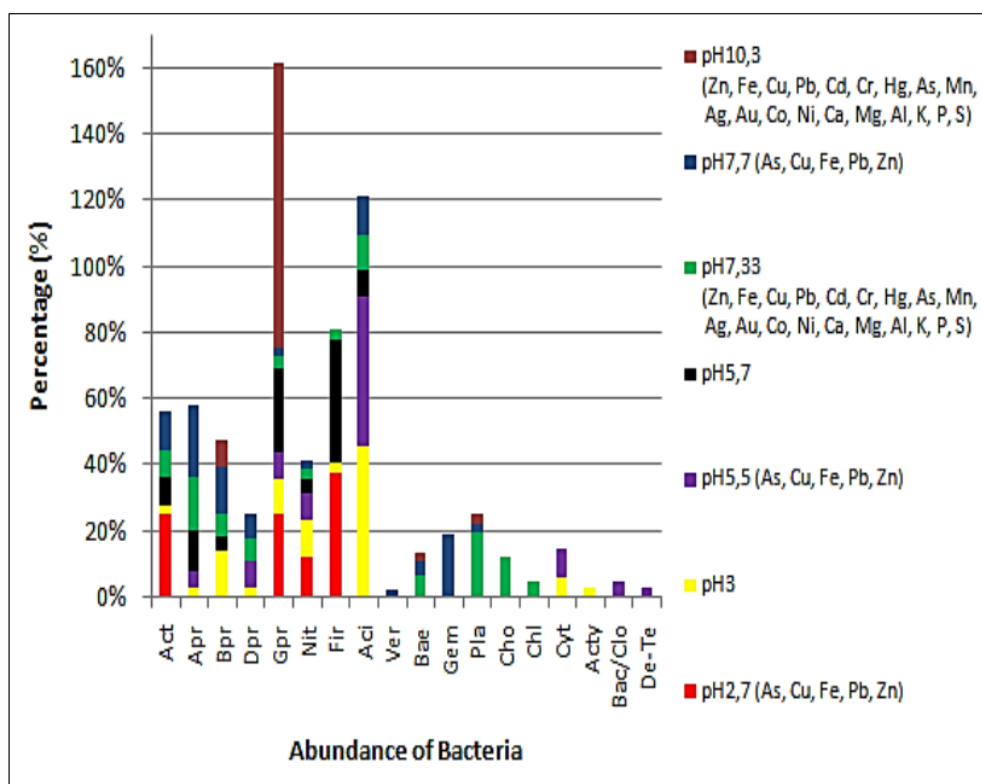


Fig. 1: Abundance Relative of Bacteria Communities in Relation to pH Value and Heavy Metal. Act (Actinobacteria), Apr (Alpha-Proteobacteria), Bpr (Beta-Proteobacteria), Dpr (Delta-Proteobacteria), Gpr (Gamma-Proteobacteria), Nit (Nitrospira), Fir (Firmicutes), Aci (Acidobacteria), Ver (Verrucomicrobia), Bae (Bacteroidetes), Gem (Gemmatimonadetes), Pla (Planctomycetes), Cho (Choloflexi), Chl (Chlorobi), Cyt (Cytophagales), Acty (Actinomycetes), Bac/Clo (Bacillus/Clostridium), De-Te (Deinococcus-Termus) [43, 49,50].

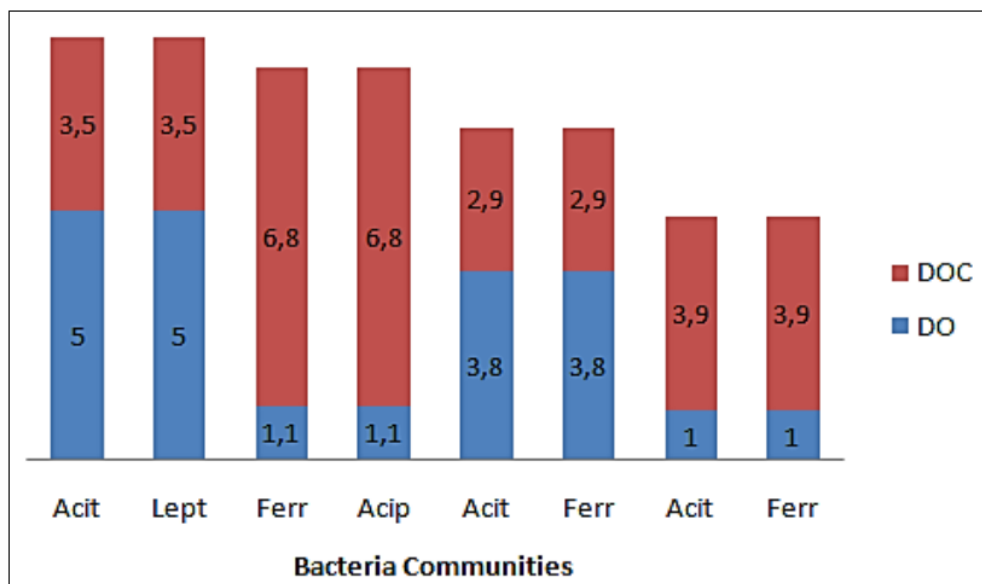


Fig. 2: The Dominant Bacteria Communities in Different DO and DOC, Based on 16S rRNA. Acit (Acidithiobacillus), Lept (Leptospirillum), Ferr (Ferroplasma), Acip (Acidiphilium) [53].

Effect of Heavy Metal on Microorganism Communities

Beside the pH and DO value, heavy metals are one of barrier factors for the growth of microorganisms. Metal concentration of water and soil are good indicators of degree of contamination [31]. The number of microorganisms depends on the total content and concentrations of particular form of heavy metals. Heavy metals can shift the structure of microbial populations, impoverish their diversity, and affect species compositions, reproduction, and activity of indigenous microorganisms [54].

Pollution has a significant impact on structural and functional diversity of microbial communities [55], size and activity [56], composition of microbial communities, and diversity change along chemical gradients [57]. A significant correlation was found between the microorganism communities and geochemical factors, including heavy metal [58]. Therefore, only certain microorganisms can survive, tolerate, or resist heavy metal contamination by determining cellular responses and adaptive mechanisms [59]. The type and composition of heavy metals can drive a diversity of microorganisms [43, 49,

50]. The impact of heavy metal contamination on microorganism (e.g. bacteria) community is shown in Figure 1. Other studies also represent the effect of heavy metal on microorganism communities diversity [53, 60, 61], and the community structure change may also occur in post tin mining pond with low oxygen concentration. Although, a study is required for a justification.

These characteristics and metabolic abilities were used to adapt in contaminated ecosystem. The changes in microbial community structure were significantly related to the size of microbial biomass as well as numerous edaphic variables (including pH and C, N, and P nutrient concentrations) [62]. Microorganism communities can survive for a long ecosystem evolution because they have resistance (insensitivity to disturbance) and resilience (the rate of recovery after disturbance) [63] for C, N, sulfur cycling, and metal [64] higher than others.

They can also use some metals as micronutrients and for redox processes, to stabilize molecules through electrostatic interactions, as components of various enzymes, and for regulation of osmotic

pressure. They remove the heavy metals by using chemicals for their growth and development. They are capable of dissolving metals and reducing or oxidizing transition metals. They have adapted to the presence of both nutrient and nonessential metals by developing a variety of biogeochemical mechanisms, such as exclusion by permeability barrier, intra- and extracellular sequestration, active transport efflux pumps, enzymatic detoxification, and reduction in the sensitivity of cellular targets to metal ions, translocation, transformation, chelation, immobilization, solubilization, precipitation, volatilization, and complexation of heavy metals [65–67].

CONCLUSIONS

Ecological changes such as presence of heavy metals, pH value, DO, and age of post mining ponds factors as a result of tin mining, particularly, in disused or post tin mining ponds predispose the structure of microorganism communities. Some studies have explained characteristics of post tin mining ponds and the microorganisms in disused tin mining ponds, including MOB and AOB. However, the survival capability of microorganisms, resistance and resilience capacity, and the interaction of microorganisms with ecosystem changes chronosequence has not been explained properly.

This review has collected information on ecological changes in disused ponds of post tin mining and response of microorganism communities on the changes chronosequencely. Further, advanced research is required about the effect of pH, DO, and heavy metals on abundance and diversity of microorganisms in post tin mining ponds, and also the consortium affectivity of bioremediator to remediate heavy metals in post tin mining pond so that it becomes a solution to accelerate the recovery of water sources for secondary activities and habitable for organisms.

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